

Effects of Lead-Contaminated Tropical Garden Soil on Bioconcentration and Physiology of the Springtail *Cyphoderus Javanus* (Börner, 1906)

PRIYANKA SARANGI¹, PARTHA PRATIM CHAKRAVORTY^{1*} and BHABATOSH DAS²

¹PG Department of Zoology, Raja Narendralal Khan Women's College (Autonomous), Affiliated to Vidyasagar University, West Bengal, India.

²Translational Health Science and Technology Institute, NCR Biotech Cluster, Faridabad, Gurugram, India.

Abstract

Lead contamination in soil represents a serious threat to soil-dwelling invertebrates. Despite the ecological importance of Collembola, their responses to heavy metal stress in tropical environments remain insufficiently explored. This study investigated the physiological stress response and lead accumulation in the soil microarthropod *Cyphoderus javanus* under chronic exposure to sublethal concentrations of lead-contaminated garden soil under tropical conditions. Individuals were maintained in an environmental chamber at 28 ± 0.5 °C, with soil moisture regulated using distilled water. Lead toxicity was evaluated by assessing moulting frequency, reproductive output, survival, and the bioconcentration factor (BCF). The results demonstrated concentration-dependent physiological stress, evidenced by reduced moulting frequency, decreased reproductive output, and lower survival at increased lead concentrations. BCF analysis indicated a time- and concentration-dependent increase in lead accumulation, although the relationship was not strictly linear, particularly at lower exposure levels. Nevertheless, lead bioaccumulation was significantly elevated with prolonged exposure and higher concentrations. These findings highlight the sensitivity of soil Collembola to lead contamination and underscore the ecological risks posed by heavy metal pollution. The study emphasizes the need to protect soil fauna to maintain ecosystem stability and soil health, which are essential for sustainable terrestrial ecosystem functioning.



Article History

Received: 27 February 2026
Accepted: 28 April 2026


Keywords

Bioindicator;
Bioconcentration;
Collembola;
Ecotoxicology;
Lead Toxicity.

CONTACT Partha Pratim Chakravorty ✉ parthapratimchakravorty@yahoo.in 📍 PG Department of Zoology, Raja Narendralal Khan Women's College (Autonomous), Affiliated to Vidyasagar University, West Bengal, India.



© 2026 The Author(s). Published by Enviro Research Publishers.

This is an  Open Access article licensed under a Creative Commons license: Attribution 4.0 International (CC-BY).

Doi: <http://dx.doi.org/10.12944/CWE.21.1.14>

Introduction

Heavy metal pollution represents a serious threat to soil ecosystems, affecting soil quality, biodiversity, and the functioning of soil-dwelling organisms. Metals enter the biosphere via natural processes such as volcanic activity and weathering, as well as through anthropogenic activities like mining, fuel combustion, and various industrial operations.^{1,2} Lead (Pb)-contaminated soils have become a significant concern, drawing growing attention from national and international regulatory bodies.^{3,4} Pb, primarily introduced into the soil through industrial activities and mining, is a persistent contaminant known for its toxic effects on soil ecosystems. Its accumulation disrupts soil biodiversity and affects the functioning of soil organisms essential for nutrient cycling and ecosystem health.^{5,6} Additionally, it may impair the growth and reproductive function of soil fauna, including earthworms, mites, as well as springtails leading to reduced egg laying, poor hatching success, decreased survival and, weight reduction.⁴

Soil provides essential habitat for diverse organisms, many of which are directly or indirectly involved in key processes like decomposition and soil nutrient process. Within the soil-associated arthropods, collembola constitutes a highly abundant group with diverse feeding habits and wide spatial distribution.⁷⁻⁹ The remarkable diversity of collembola is driven by their varied life forms and specialized ecological traits. Relative to natural habitats, the agricultural environments show a significant reduction of richness and population abundance of collembola.¹⁰ This decline has been attributed to elevated pollutant levels and limited resource availability.⁷ The reduction of collembolan populations in agricultural systems can adversely impact nutrient cycling, ultimately leading to decreased agricultural productivity.^{11,8}

Heavy metal pollution is known to negatively impact the growth of Collembola, making them a useful indicator for measuring ecological toxicity. For example, studies have linked arsenic-contaminated soils to reduced body size in Collembola, identifying growth retardation as an initial indicator of environmental pollution.^{12,13} Similarly, cadmium exposure has been found to inhibit growth even at low concentrations, with growth emerging as a more sensitive indicator than reproduction in some cases.¹³ Metals such as cadmium and lead have consistently demonstrated a significant suppressive

effect on Collembola growth.^{13,14} Research in ecotoxicology frequently targets two key species, *Folsomia candida* (parthenogenetic) and *Folsomia fimetaria* (sexual), due to their well-characterized biology and suitability for laboratory experiments.¹³ However, while the impacts of heavy metals on Collembola have been extensively studied, most of this work has been limited to European soils,^{10,15-18} leaving a critical gap in understanding how these organisms respond to contamination in tropical soil environments. In this context, *Cyphoderus javanus*, a soil-dwelling collembola commonly found in tropical environments, was selected as a model organism due to its ecological relevance and potential sensitivity to soil contaminants, making it suitable for assessing region-specific ecotoxicological risks.¹⁹ This research sought to investigate the influence of Pb-contaminated soil on organismal physiology of *C. javanus* and to assess its capacity for Pb accumulation through bioconcentration factor (BCF) analysis. Given the persistence of Pb in soils and its potential risks to soil-dwelling organisms, this research provides insights into the ecological impacts of heavy metal contamination. The results are anticipated to contribute to the development of soil quality guidelines and regulatory frameworks for tropical regions.

Materials and Methods

Experimental Organism

C. javanus was collected from organic-rich soil (sampling site free from agricultural disturbances) and cultured in 4 cm × 5.5 cm polythene vials containing 7 g of moist garden soil.¹⁹ The organisms were kept at 28°C ± 0.5°C in an environmental chamber, with baker's yeast (2 mg per vial) used as food and water added to maintain the moisture at every 2–3 days to keep the substrate near its water-holding capacity.²⁰ Juveniles were reared separately after being transferred from the adult vials. Toxicity tests were conducted using 12- to 15-day-old juveniles.

Exposure Medium

Soil obtained from a natural garden area which is free from any agricultural activities was used as the testing medium for the bioassay experiments. The collected soil was air-dried, homogenized, and sieved to remove debris prior to use. Key physicochemical properties¹⁸ (Table 1) of the test medium - soil texture,²¹ pH, organic carbon²² and,

water holding capacity^{23,24} were determined. Prior to the bioassay, the soil was defaunated through heat treatment at 70 °C for 2 h, and then stored in the dark at 5 °C.²⁵

Table 1: Physicochemical characteristics of the tested soil¹⁸

Properties of soil	Observed value
Soil Texture (g %)]	
Clay	17.25
Silt	11.75
Sand	71.00
Soil pH	6.80
Organic Carbon Content of the Soil	1.45%
Soil water retention capacity	34.40%

Experimental Chemical

The inadequate disposal of fly ash, a byproduct of coal combustion, presents a major environmental concern, especially in developing countries.²⁶ Fly ash, commonly utilized in agriculture for its alkaline properties and mineral content that enhance soil health, poses an environmental risk due to its increasing levels of hazardous heavy metals such as arsenic, lead, cadmium, chromium, and mercury.²⁷ Notably, the heavy metal burden in fly ash is 2-10 times that of the source coal.²⁸ In this study, soil contaminated with fly ash was obtained from the Kolaghat Thermal Power Plant in West Bengal, India (22°25'28.3"N, 87°51'39.4"E), following a randomized block layout to ensure a structured sampling method. The heavy metal content was quantified using Atomic Absorption Spectroscopy (AAS), and lead, the predominant metal, was chosen for further investigation. Lead acetate trihydrate [(CH₃COO)₂Pb·3H₂O], sourced from Merck Life Science, was employed in the experimental procedures.

Acute Toxicity Test

The study investigated the short-term toxic effects of lead acetate on *C. javanus*. Groups of ten 12- to 15-day-old *C. javanus* juveniles were exposed to various concentrations of lead acetate, mimicking lead levels found in fly ash-polluted soil. Each concentration had three replicates. The lead solutions were mixed uniformly into the soil before adding the organisms. After 24 hours, the number of

dead *C. javanus* was recorded to determine the lead concentration causing 50% mortality (LC50). The experimental procedures were based on standard protocols but included some modifications.²⁹

Chronic Toxicity Test on the Physiology of *C. Javanus*

Following standard procedures,³⁰ this study examined how lead acetate affects the shedding of exoskeletons (exuvia production), the egg laying capacity (fecundity), and the lifespan of *C. javanus*. Based on the previous acute toxicity test, sublethal lead acetate concentrations in the soil: 5933.5, 2966.75, 1977.83, and 1483.37 mg/kg has been chosen. 45 small plastic vials (4 cm wide, 5 cm tall), each with 7 grams of soil mixed with one of the lead acetate concentrations has been prepared. Each concentration was tested in three identical vials, and ten 12- to 15-day-old juvenile *C. javanus* were placed in each vial. The number of exuviae and eggs laid per individual were recorded every 24 hours over a 28-day period.

Analysis of BCF in *C. Javanus*

Lead (Pb) concentrations in *C. javanus* were measured to evaluate its bioconcentration potential in contaminated soil. To ensure sufficient biological material for metal analysis, *C. javanus* specimens were pooled into composite samples. The dry weight of each sample was analyzed using a micro-analytical balance (Mettler Toledo).

Acid digestion was performed using concentrated nitric acid (HNO₃) to mineralize the samples. Approximately each sample was treated with 1–2 mL of concentrated HNO₃ and digested at 80–100 °C until complete digestion was achieved. Following digestion, the samples were allowed to cool and subsequently diluted with distilled water to the desired volume prior to analysis.³¹ The resulting solution was analyzed using Atomic Absorption Spectroscopy (AAS) as per IS: 3025 (Part 47):1994, employing an air-acetylene flame and a hollow-cathode lamp at 283.3 nm.³² Standard lead solutions were prepared from certified reference material (CRM), and a calibration curve was used for quantification. Analytical accuracy was ensured using reagent blanks and quality controls. Lead concentrations were used to calculate the BCF, reflecting the ability of the species to accumulate lead from the environment. The BCF was determined

by the ratio of Pb concentration in the collembola by the total Pb concentration in the soil.^{33,34}

Data Processing and Statistical Analysis

The data were computed and visualized using SPSS software (SPSS v.29.0) and Microsoft Excel. To estimate the LC₅₀ value—the concentration of lead causing 50% mortality in springtails—probit analysis has been conducted using SPSS. Results are presented as mean values with corresponding standard deviations (SD) to indicate variability. To assess statistically significant differences in the responses of springtails across varying lead

exposure levels, one-way ANOVA was conducted. Upon identification of significant difference, Tukey's post hoc test was used to indicate which treatment groups differed significantly. A significance threshold of $p < 0.05$ was adopted throughout the statistical evaluation.

Results

Short-term Toxic Effects of lead Acetate on *C. Javanus*

Following the 24-hour of exposure the LC₅₀ value of lead acetate for *C. javanus* is 11,867.13 mg/kg (Table 2).

Table 2: 24-hour median lethal concentration (LC₅₀) values of lead acetate for *C. javanus*¹⁷

Metallic Pollutant	24 hours median lethal concentration	95% confidence Interval (CI)
Lead Acetate	11,867.13 mg/kg	7806.334-14773.270

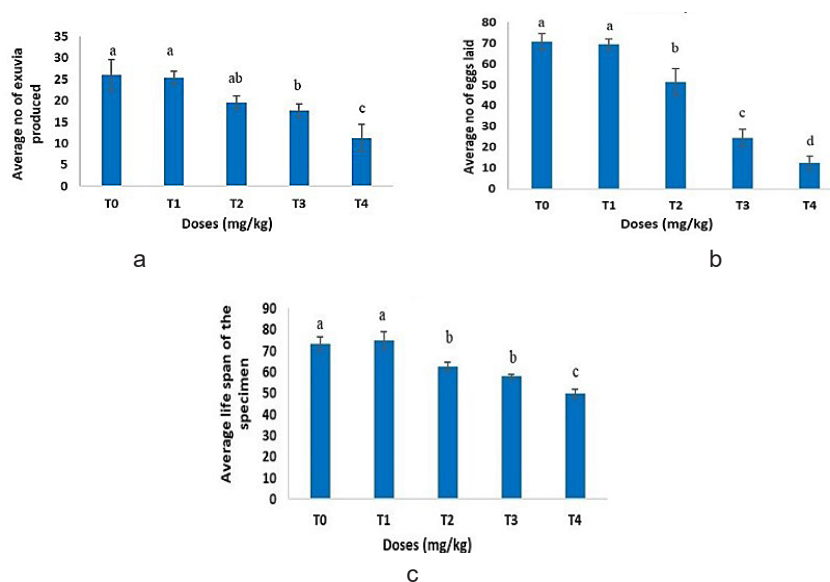


Fig. 1: Exuvia production (a), fecundity (b), and lifespan (c) of *Cyphoderus javanus* under sublethal Pb concentrations: T1 (1/8 of the LC₅₀), T2 (1/6 of the LC₅₀), T3 (1/4 of the LC₅₀), and T4 (1/2 of the LC₅₀), along with the control group (T0). Each treatment consisted of three replicate vials, with 10 individuals per vial ($n = 3$). Data are presented as mean \pm standard deviation (SD).

Statistically significant differences ($p < 0.05$) among treatments are denoted by different letters above the bars.

Chronic Effects of Lead Acetate on *C. Javanus*

During the lead exposure period, clear long-term toxic effects were evident. The control group exhibited an average molting frequency of (26 ± 3.606). In contrast, molting declined significantly ($p < 0.05$) in the T3 group (17.67 ± 1.528) and the T4 group (11.33 ± 3.215), corresponding to 1/4 and 1/2 of the LC50 concentrations, respectively as reflected by distinct letter groupings (Fig. 1a). On the other hand, the T1 (1/8 LC50) and T2 (1/6 LC50) groups showed non-significant difference in molting compared with the control, sharing similar letter groupings.

Reproductive capacity (fecundity) also significantly decreased ($p < 0.05$) as lead concentrations increased. The control group had a mean fecundity of 70.67 ± 3.786 , whereas significant reductions were found in the T2 (51.33 ± 6.506), T3 (24.33 ± 4.163), and T4 (12.33 ± 3.512) groups (Fig 1b).

Lead exposure also impacted the lifespan of *C. javanus*. The average lifespan of the control group was 73.33 ± 3.21 days, with significant reductions ($p < 0.05$) observed in the T2, T3, and T4 treatments. However, the T1 group did not show a significant change in lifespan (Fig1c).

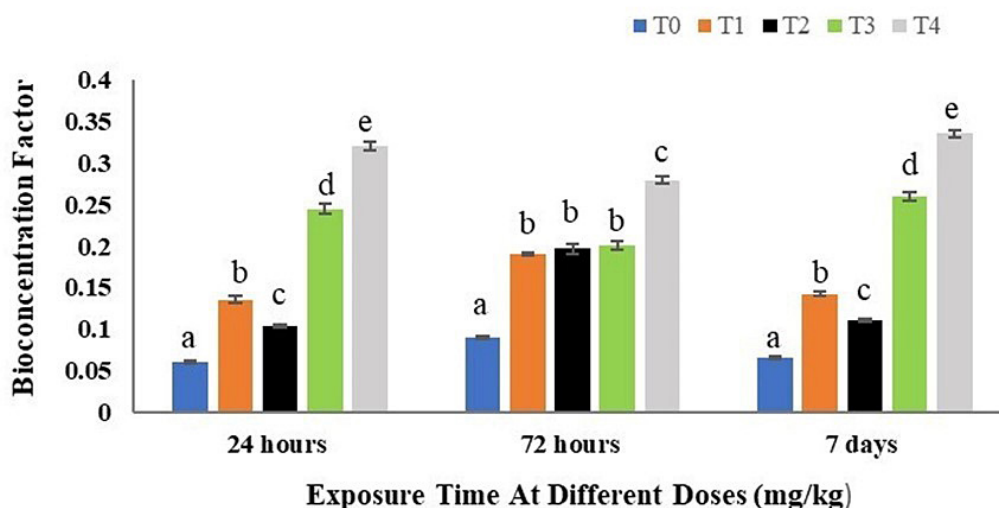


Fig. 2: Bioconcentration factor (BCF) of Pb in *Cyphoderus javanus* after 24 h, 72 h, and 7 d of exposure to sublethal concentrations: T1 (1/8 of the LC₅₀), T2 (1/6 of the LC₅₀), T3 (1/4 of the LC₅₀), and T4 (1/2 of the LC₅₀), along with the control group (T0). BCF was computed as the ratio between Pb concentration in the organism to that in the surrounding soil. For each replicate (n = 3), pooled samples consisting of 50 individuals were used for analysis. Data are illustrated as mean values with standard deviation (SD).

The bioconcentration factor (BCF) of *C. javanus* exhibited a general increase with rising lead concentrations and longer exposure durations. At 24 hours (Fig. 2), BCF values varied between 0.060 ± 0.002 in the control group to 0.320 ± 0.005 at T4. One-way ANOVA indicated significant difference among treatments ($F_{4, 10} = 2040.01$, $p < 0.001$). A significant increase ($p < 0.05$) in BCF was noted across all treatments compared to the control, with each dose level differing significantly from

the others. However, BCF at T2 (0.104 ± 0.002) was significantly lower than at T1 (0.136 ± 0.004), indicating a non-linear accumulation pattern at lower doses despite overall dose-dependent trends. At 72 hours, the control BCF was 0.090 ± 0.002 , while the highest value was recorded at T4 (0.279 ± 0.004). ANOVA indicated significant differences among treatments ($F_{4, 10} = 787.46$, $p < 0.001$). While T3 (0.201 ± 0.005) and T2 (0.197 ± 0.006) had similar mean values, statistical tests showed that they were

not significantly differed from each other ($p > 0.05$). In contrast, both were significantly higher than T1 (0.190 ± 0.003) and the control ($p < 0.05$), reinforcing a partial dose-dependent response with a plateau between T2 and T3. By day 7, BCF values peaked at 0.336 ± 0.004 in the T4 group, compared to 0.065 ± 0.002 in the control. ANOVA again showed significant variation among treatments ($F_{4, 10} = 3377.84$, $p < 0.001$). The values for T1 (0.142 ± 0.003) and T2 (0.110 ± 0.002) revealed a significant difference ($p < 0.05$), confirming that the earlier deviation in accumulation trends was also evident at this later stage. From T3 (0.260 ± 0.005) onward, a clear dose- and time-dependent increase in BCF was observed, with all treatments significantly different from one another and from the control.

These findings demonstrate that *C. javanus* accumulates Pb in a generally dose- and time-responsive manner; however, accumulation was not strictly linear at lower concentrations, possibly reflecting physiological thresholds or regulatory responses limiting uptake at mild exposure levels.

Discussion

Biological Consequences of Lead Contamination on *C. Javanus*

This study found that exposure to Pb severely damages important biological activities in *C. javanus*. A dose-dependent reduction in exuvia production, fecundity, and overall lifespan has been documented especially at high Pb concentrations. Shedding of the exoskeleton requires substantial energy and is frequently associated with increased mortality in cultured organisms.³⁵ These findings are consistent with recent research on the collembola *F. candida*, which likewise showed significant decreases in reproductive output and adult survival after Pb exposure.¹³ Similarly, investigations conducted in urban soils with high levels of Pb and cadmium (Cd) found significant detrimental effects on the reproduction, survival, and development of both *F. candida* and *Ceratophysella caeca*.^{4,6,16}

While some investigations report that the survival of *F. candida* is not substantially affected by Pb,³⁴ its reproductive performance is consistently shown to be a more sensitive endpoint. For example, *F. candida* reproduction was significantly reduced in spiked soils with an overall Pb content of 3,877 mg/

kg.³⁶ Similarly, significant reproductive impairments were reported at the highest Pb concentrations (range from 0 to 600, 1200, 2400, and 4,800 mg/kg) across three soil types.³¹ The current work reveals that *C. javanus* is similarly sensitive to Pb toxicity, supporting the theory that important reproductive and developmental parameters are common vulnerability areas for collembola under heavy metal stress. However, there is interspecific variation in metal tolerance; for example, the tropical species *Xenylla welchi* has exhibited even greater susceptibility, particularly following long-term lead exposure.¹⁸ Such variances highlight the need of including multiple species in ecotoxicological assessments to gain a more complete picture of the ecological dangers posed by soil contamination.

Bioaccumulation of Lead in *C. Javanus*

The computed BCF for Pb in *C. javanus* reveals a distinct pattern of metal buildup that is depending on both the contaminant concentration and the length of exposure. At each evaluated time point (24 hours, 72 hours, and 7 days), BCF levels tended to rise as lead concentrations in the garden soil increased. This reveals that *C. javanus* actively absorbs lead from its surroundings, with internal concentrations positively correlated with both exterior lead levels and exposure duration.³⁷ These findings are comparable with prior findings for Pb-contaminated soils^{32,34,38} and with accumulation patterns found for other hazardous elements such as Cd³⁹⁻⁴³ and copper (Cu).⁴

The observed bioaccumulation of lead is directly related to the physiological stress that *C. javanus* exhibits, which includes shortened lifespan, decreased ability to reproduce, and poor molting. The relationship between higher BCF values and more severe physiological impairments implies that the harmful effects of Pb on essential biological systems are mostly caused by internal accumulation.^{36,44} To handle this accumulated lead, *C. javanus* will probably need to expend a significant amount of energy. Detoxification, cellular repair, and homeostasis may get resources that would otherwise be used for growth, reproduction, and survival. This need for energy to power defensive mechanisms probably gets stronger as BCF values increase with increased lead concentrations and longer exposure times.⁴⁵

The study emphasizes that when evaluating the ecological risk posed by heavy metal contamination for soil invertebrates, it is essential to incorporate both physiological endpoints and bioaccumulation data. Measurable buildup of Pb in *C. javanus* and the ensuing negative physiological impacts show that bioaccumulation can be a powerful indicator of possible ecological damage. Therefore, a comprehensive risk assessment must take into account both the extent of pollutant intake and its eventual effects on the fitness and health of the organisms.⁴⁶

Conclusion

This study demonstrates that *C. javanus* accumulates Pb in a concentration- and time-dependent manner, as reflected by increasing BCF values. The associated physiological impairments indicate that internal metal accumulation exerts significant toxic effects. These findings suggest that Pb contamination in garden soils can adversely affect soil microarthropods, potentially disrupting soil health and ecosystem stability. Compared to commonly studied species such as *F. candida*, *C. javanus* exhibited comparable sensitivity under the tested conditions, supporting its suitability as a bioindicator in tropical soils. Integrating bioaccumulation data with physiological responses provides a more comprehensive structured approach for assessing ecological risk. These results may aid in the development of region-based soil quality guidelines and management strategies for mitigating heavy metal contamination in tropical environments.

Acknowledgement

The authors gratefully acknowledge the facilities provided by the Principal and the Director of the Natural and Applied Sciences Research Centre, Raja Narendra Lal Khan Women's College (Autonomous),

Paschim Medinipur, West Bengal, India. PS acknowledges the financial support received in the form of Junior Research Fellowship (JRF) and Senior Research Fellowship (SRF) from the University Grants Commission (UGC) for conducting this research work, and PPC acknowledges the Department of Science and Technology–CURIE (DST-CURIE) for infrastructure support.

Funding Sources

UGC-India, JRF & SRF. Reference no-221610-157814.

Conflict of Interest

The authors do not have any conflict of interest.

Data Availability Statement

This statement does not apply to this article.

Ethics Statement

This research did not involve human participants, animal subjects, or any material that requires ethical approval.

Informed Consent Statement

This study did not involve human participants, and therefore, informed consent was not required.

Permission to reproduce material from other sources

Not Applicable

Author Contributions

- **Priyanka Sarangi:** Writing Original Draft, Statistical Analysis.
- **Partha Pratim Chakravorty:** Supervision, Review and Editing.
- **Bhabatosh Das:** Supervision.

References

1. Garrett RG. Natural Sources of Metals to the Environment. *Hum. Ecol. Risk Assess.* 2000; 6(6), 945–963. <https://doi.org/10.1080/10807030091124383>.
2. Dobritzsch D, Grancharov K, Hermsen C, Krauss G-J, Schaumlöffel D. Inhibitory effect of metals on animal and plant glutathione transferases. *J Trace Elem Med Biol.* [online]. 2020; 57, 48–56. <https://doi.org/10.1016/j.jtemb.2019.09.007>.
3. Migliorini M, Pigo G, Bianchi N, Bernini F., Leonzio, C. The effects of heavy metal contamination on the soil arthropod community of a shooting range. *Environ Pollut.* 2004; 129: 331–340.
4. Ding Y, Li Z, Ke X, Wu L, Zuo S. Toxicity of lead

- pollution to the collembolan *Folsomia candida* in Ferri-Udic Cambosols. *Pedosphere*. 2021; 31, 627–637.
5. Callahan ST, Bidwell A, Lin C, DeLuca TH, Tobin PC. Effects of copper exposure and increased temperatures on *Collembola* in western Washington, USA. *City Environ. Interact.* [online]. 2019; 4, 100026. <https://doi.org/10.1016/j.cacint.2020.100026>.
 6. Dai W, Holmstrup M, Stine Slotsbo Z, Ke X, Li Z, Gao M, Wu L. Compartmentation and effects of lead (Pb) in the collembolan, *Folsomia candida*. *Environ Sci Pollut Res*. 2020; 27(35), 43638–43645. <https://doi.org/10.1007/s11356-020-10300-6>.
 7. Sterzyńska M, Nicia P, Zandrozny P, Fiera C, Shrobovych J. Urban springtail species richness decreases with increasing air pollution. *Ecol Indic*. 2018; 94:328–335. <https://doi.org/10.1016/j.ecolind.2018.06.063>.
 8. Bardgett RD, van der Putten WH. Belowground biodiversity and ecosystem functioning. *Nat*. 2014;515(7528):505–511. <https://doi.org/10.1038/nature13855>
 9. Sun X, Marian F, Bluhm C, Maraun M, Scheu S. Response of *Collembola* to the addition of nutrients along an altitudinal gradient of tropical montane rainforests. *Appl Soil Ecol*. 2020; 147:103382. <https://doi.org/10.1016/j.apsoil.2019.103382>.
 10. Syrek D, Weiner WM, Wojtylak M, Olszowska G, Kwapis Z. Species abundance distribution of collembolan communities in forest soils polluted with heavy metals. *Appl Soil Ecol*. 2006; 31:239–250. <https://doi.org/10.1016/j.apsoil.2005.05.002>.
 11. Jernigan A, Kao-Kniffin J, Pethybridge S, Wickings K. Soil microarthropod effects on plant growth and development. *Plant Soil*. 2023 Feb;483(1):27-45.
 12. Gruss I, Twardowski J, Karczewska A, Szopka K, Kluczek K, Magiera-Dulewicz J. *Collembola* reduce their body sizes under arsenic contamination in the soil – possible use of new screening tool in ecotoxicology. *Ecol Indic*. 2022; 142, 109185. <https://doi.org/10.1016/j.ecolind.2022.109185>
 13. Gruss I, Lallaouna R, Twardowski J, Magiera-Dulewicz J, Twardowska K. *Collembola* growth in heavy metal-contaminated soils. *Sci rep.* [online], 2024; 14(1), p.27998. <https://doi.org/10.1038/s41598-024-79766-5>.
 14. Fountain MT, Hopkin SP. Continuous Monitoring of *Folsomia candida* (Insecta: Collembola) in a Metal Exposure Test. 2001; 48(3), 275–286. <https://doi.org/10.1006/eesa.2000.2007>.
 15. Liu M, Xu J, Krogh PH, Song J, Wu L, Luo Y, Ke X. Assessment of toxicity of heavy metal-contaminated soils toward *Collembola* in the paddy fields supported by laboratory tests. *Environ Sci Pollut Res*. 2018; 25(17), pp.16969–16978. <https://doi.org/10.1007/s11356-018-1864-y>.
 16. Kayiranga A, Li Z, Isabwe A, Ke X, Simbi CH, Ifon BE, *et al.* The effects of heavy metal pollution on collembola in urban soils and associated recovery using biochar remediation: A review. *Int. J. Environ. Res. Public Health*. 2023; 20(4):3077. <https://doi.org/10.3390/ijerph20043077>
 17. Sarangi P, Chakravorty PP, Das B. Comprehensive analysis of biochemical responses in a soil collembola (*Cyphoderus javanus* borner, 1906) exposed to lead-contaminated garden soil: Implications for soil health and ecological risk assessment. *Int. J. Entomol. Res*. 2024; 9(10), 58-65.
 18. Sarangi P, Chakravorty PP, Das B. The Usefulness of physiological and biochemical stress response of soil collembola (*Xenylla welchi* Folsom, 1916) as a biomarker in a lead-amended garden soil. *Environ Anal Health Toxicol*. 2025; 40(1), e2025004. <https://doi.org/10.5620/eaht.2025004>.
 19. Sahana A, Agarwal S, Bhattacharya S, Chacko JV. Short-term oxidative stress responses in *Cyphoderus javanus* Borner (*Collembola*), as biomarkers of heavy metal pollution in lateritic soil. *Pollut Res*. 2014; 33(4):201-206.
 20. Chakravorty PP, Haque A, Sanyal S, Dasgupta R. Effect of herbicides on *Cyphoderus javanus* (Hexapoda: *Collembola*) under laboratory conditions. *J Entomol Zool Stud*. 2015; 3(1):220-223.
 21. Piper CS. Soil and plant analysis. *Hassell Press*; 1942.
 22. Walkey A, Black IA. An examination of the Degtjareff method for determining soil organic matter, and a proposed modification of the chromic acid titration method. *Soil Sci*.

- 1934; 37(1):29-38. <https://doi.org/10.1097/00010694-193401000-00003>.
23. Lal R. Physical properties and moisture retention characteristics of some Nigerian soils. *Geoderma*, 1978, 21(3):209-223. [https://doi.org/10.1016/0016-7061\(78\)90028-9](https://doi.org/10.1016/0016-7061(78)90028-9)
24. Viji R, Rajesh PP. Assessment of water holding capacity of major soil series of Lalgudi, Trichy, India. *J. Environ. Res. Dev.* 2012; 7(1A):393-398.
25. Wiles JA, Frampton GK. A field bioassay approach to assess the toxicity of insecticide residue on soil to Collembola. *Pest Manag Sci.* 1996; 47(3):273-285.
26. Panda RB, Biswal T. Impact of fly ash on soil properties and productivity. *Int. j. agric. environ. Biotechnol.* 2018; 11(2):275-283. <https://doi.org/10.30954/0974-1712.04.2018.8>.
27. Jambhulkar HP, Shaikh SMS, Kumar MS. Fly ash toxicity, emerging issues and possible implications for its exploitation in agriculture; Indian scenario: A review. *Chemosphere.* 2018; 213, 333–344. <https://doi.org/10.1016/j.chemosphere.2018.09.045>.
28. Chen Y, Fan Y, Huang Y, Liao X, Xu W, Zhang T. A comprehensive review of toxicity of coal fly ash and its leachate in the ecosystem. *Ecotoxicol Environ Saf.* [online], 2024; 269, p.115905. <https://doi.org/10.1016/j.ecoenv.2023.115905>.
29. International Organization for Standardization (ISO). Soil quality - Inhibition of reproduction of Collembola (*Folsomia candida*) by soil pollutants. [cited Nov 6, 2024]. Available from: <https://www.iso.org/standard/19245.html>.
30. Organisation for Economic Co-operation and Development (OECD). Test No. 232: *Collembolan Reproduction Test in Soil*. [cited Nov 6, 2024]. Available from: https://www.oecd.org/en/publications/test-no-232-collembolan-reproduction-test-in-soil_9789264264601-en.html
31. US EPA. Acid digestion of sediments, sludges, and soils (Method 3050B). *Washington, DC*; 1996.
32. Bur T, Crouau Y, Bianco A, Gandois L, Probst A. Toxicity of Pb and of Pb/Cd combination on the springtail *Folsomia candida* in natural soils: Reproduction, growth and bioaccumulation as indicators. *Sci Total Environ.* 2012; 414, 187-197.
33. Mountouris A, Voutsas E, Tassios D. Bioconcentration of heavy metals in aquatic environments: the importance of bioavailability. *Mar Pollut Bul.* 2002; 44(10), 1136–1141, 2002. [https://doi.org/10.1016/S0025-326X\(02\)00168-6](https://doi.org/10.1016/S0025-326X(02)00168-6).
34. Luo W, Verweij RA, van Gestel CAM. Assessment of the bioavailability and toxicity of lead polluted soils using a combination of chemical approaches and bioassays with the Collembolan *Folsomia Candida*. *J Hazard Mater.* 2014a; 280: 524–530.
35. Kristiansen SM, Borgå K, Rundberget JT, Leinaas HP. Effects on life-history traits of *Hypogastrura viatica* (Collembola) exposed to imidacloprid through soil or diet. *Environ Toxicol Chem.* 2021 Nov 1;40(11):3111-22.
36. Lock K, Waegeneers N, Smolders E, Criel P, Van Eeckhout H, Janssen CR. Effect of leaching and aging on the bioavailability of lead to the springtail *Folsomia candida*. *Environ Toxicol Chem.* 2006; 25(8), 2006–2010.
37. Lanno RP, Oorts K, Smolders E, Albanese K, Chowdhury MJ. Effects of soil properties on the toxicity and bioaccumulation of lead in soil invertebrates. *Environ Toxicol Chem.* 2019; 38(7), 1486–1494.
38. Bongers M, Rusch B, van Gestel CAM. The effect of counterion and percolation on the toxicity of lead for the springtail *Folsomia candida* in soil. *Environ Toxicol Chem.* 2004; 23: 195–199.
39. Crommentuijn T, Brils J, Vanstraelen NM. Influence of cadmium on life-history characteristics of *Folsomia candida* (Willem) in an artificial soil substrate. *Ecotoxicol Environ Saf.* 1993; 26: 216–227.
40. van Gestel CAM, Hensbergen PJ. Interaction of Cd and Zn toxicity for *Folsomia candida* Willem (Collembola: Isotomidae) in relation to bioavailability in soil. *Environ Toxicol Chem.* 1997; 16: 1177–1186.
41. Ardestani MM, Ortiz MD, van Gestel CAM. Influence of Ca and pH on the uptake and effects of Cd in *Folsomia candida* exposed to simplified soil solutions. *Environ Toxicol Chem.* 2013; 32: 1759–1767.

42. Zhu D, Ke X, Wu LH, Christie P, Luo YM. Biological transfer of dietary cadmium in relation to nitrogen transfer and ¹⁵N fractionation in a soil collembolan-predatory mite food Chain. *Soil Biol Biochem.* 2016a; 101: 207–216.
43. Zhu D, Ke X, Wu LH, Li, Z., Christie, P., Luo, Y.M. Ecotoxicity of cadmium in a soil collembolan-predatory mite food chain: Can we use the ¹⁵N labeled litter addition method to assess soil functional change? *Environ Pollut.* 2016b; 219, 37–46.
44. Luo W, Verweij R A, van Gestel CAM. Determining the bioavailability and toxicity of lead contamination to earthworms requires using a combination of physicochemical and biological methods. *Environ Pollut.* 2014c; 185, 1–9.
45. Spurgeon DJ, Hopkin SP. Seasonal variation in the abundance, biomass and biodiversity of earthworms in soils contaminated with metal emissions. *Pedobiologia.* 1999; 43(6), 641–651.
46. Lanno RP, Wells J, Conder J, Bradham K, Basta N. The bioavailability of chemicals in soil for earthworms. *Ecotoxicol Environ Saf.* 2004; 57(1), 39–47.